

Short communication

Toxicity and biodegradability of imidazolium ionic liquids

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Abstract

Several bioassays have been carried out to analyze the toxicity and biodegradability of several imidazolium ionic liquids (ILs) in aqueous phase. The synthesized compounds consist of an imidazolium cation with two alkyl substituents in positions 3 (R_1) and 1 (R_2) and a counter-ion. The alkyl substituent R_1 has been fixed as a methyl group and the effect of the alkyl chain length (C_1 – C_8) of the other substituent (R_2) has been tested. Moreover, the influence of diverse counter-ions A^- (Cl^- , PF_6^- , XSO_4^-) has been analyzed. Acute toxicity and EC_{50} values of each compound in the aqueous solution have been determined by using the Microtox[®] standard procedure. Biodegradability of IL has been determined by measuring BOD_5 of aqueous samples containing IL and/or D-glucose and the IL residual content and/or D-glucose concentration after this assay. The viability of the microorganisms used in the BOD_5 has been related to the ATP in the samples, measured by a bioluminescence assay. All the ILs tested were not biodegradable in the considered conditions. Besides, it was found that the shorter the chain length of side chain R_2 , the lower the toxic effect is. On the contrary, the anion has a little effect on the IL toxicity.

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1. Introduction

Ionic Liquids, ILs, are low-melting-point salts that have become increasingly attractive as green solvents for industrial applications [1–10]. This green adjective is mainly attributed to their negligible vapor pressure, which avoids the loss of solvent to the atmosphere and decreases the worker exposure risk. Thus, room-temperature ionic liquids could provide environmentally friendly solvents for the chemical and pharmaceutical industries [1,5,6,11]. Typical ILs consist of an organic cation with delocalized charges and a small inorganic anion, most often halogen anions weakly coordinating such as Cl^- , BF_4^- or PF_6^- [12]. An ionic liquid can be thought of as “designer” solvent [13] so it should be possible to design, or tailor, a solvent for a certain reaction. Therefore, many cation and anion combinations are possible, changing properties as polarity, hydrophobicity and solvent miscibility behavior. Among these possibilities, the 1-alkyl-3-methylimidazolium is one of the most used because it is non-volatile, non-flammable, presents high thermal stability

and is an excellent solvent for a wide range of inorganic and organic materials [6,7,9].

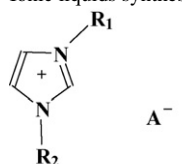
Although, the low vapor pressure of ILs may reduce the air pollution with respect to the typical volatile organic solvents, this is not enough to justify calling a “green” process. It must be considered that a release of ILs from industrial processes into aquatic environments may lead to water pollution, because of their high solubilities in water. Moreover, because of the high stability of ionic liquids in water these compounds could become as persistent pollutants in wastewaters. For this reason it is priority to determine the further consequences and the environmental risk of the presence of ILs in wastewaters.

Most employed methods to evaluate the environmental risk of a substance in an aqueous media are those measuring their toxicity by using an inhibition assay. Different microorganism or enzymes have been used in this inhibition measurements, the acute toxicity test which uses the *V. fischeri* (formerly *Photobacterium phosphoreum*) bioluminescence inhibition assay being one of the most applied [14,15]. This is a standard ecotoxicological bioassay in Europe (DIN EN ISO 11348). It is very rapid, cost-effective, and it is a widely accepted method for toxicity determination used extensively in the literature focusing on environmental issues. The experimental procedure Microtox

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Table 1
Ionic liquids synthesized and tested



$R_1 = CH_3$

Acronym	R ₂	A ⁻
[C ₁ mim][CH ₃ SO ₄]	CH ₃	CH ₃ SO ₄ ⁻
C ₂ mim][C ₂ H ₅ SO ₄]	CH ₂ -CH ₃	CH ₃ -CH ₂ SO ₄ ⁻
[C ₄ mim][Cl]	(CH ₂) ₃ -CH ₃	Cl ⁻
[C ₆ mim][Cl]	(CH ₂) ₅ -CH ₃	Cl ⁻
[C ₈ mim][Cl]	(CH ₂) ₇ -CH ₃	Cl ⁻
[C ₆ mim][PF ₆]	(CH ₂) ₅ -CH ₃	PF ₆ ⁻
[C ₈ mim][PF ₆]	(CH ₂) ₇ -CH ₃	PF ₆ ⁻

has been adopted from the official standards of several countries including USA (ASTM method D5660-1995), Germany (DIN 38412-1990), France (AFNOR T90-320-1991) and Spain (ISO 11348-3-1998).

By using the Microtox assay it has been observed an increasing toxicity trend when the alkyl chain length substituent in imidazolium ionic liquids [16–18] increases. On the other hand, minimal effects on toxicity are observed varying the anion of the methyl imidazolium salts. Consequently, toxicity of the ILs was mainly attributed to the alkyl chain.

Similar results were found using various bacteria for the inhibition test [19–22], with IPC-81 and C₆ glioma cells [17], or using higher organisms, including *Caenorhabditis elegans* and *Daphnia magna* [23,24]. Few studies have been carried out analyzing the biodegradability of ILs in the aqueous media. García et al. [16] using the closed bottle test, have measured the 28 days biochemical oxygen demand (BOD) in series of butylmethylimidazolium (bmimX, X = Br, BF₄, PF₆, NTf₂, N(CN)₂ and octylOSO₃) and methyl-(propoxycarbonyl)-imidazolium ionic liquids. They found that ILs generally proved to be poorly biodegradable. The corresponding 3-methyl-1-(propoxymethylcarbonyl)-imidazolium series showed higher levels of biodegradability but none of the compounds that were tested could be classified as “readily biodegradable”. Therefore, the ILs could become persistent pollutants and break through classical treatment systems into natural waters.

In this work, the toxicity and biodegradability of several imidazolium ILs in aqueous phase have been determined. Solutions of these compounds, summarized in Table 1, have been prepared in redistilled water and they were synthesized in our laboratory. A methyl group is at position R₁ and different alkyl chain substituents (chain length from C₁–C₈) are at position R₂. Diverse counter-ions A⁻ (Cl⁻, PF₆⁻, XSO₄⁻) have been tested. Acute toxicity and EC₅₀ values of each compound in the aqueous solution have been determined by using the Microtox[®] standard procedure. To analyze the biodegradability of IL, the biochemical oxygen demand for 5 days, BOD₅ of several aqueous samples containing known initial amounts of IL and/or

D-glucose have been determined. Furthermore, the residual IL content and/or D-glucose concentration in these aqueous samples have been measured after 5 and 10 days. Finally, the viability of the microorganisms used in the BOD₅ has been related to the APT in the samples, measured by a bioluminescence assay.

2. Experimental

ILs were prepared according to slightly modified literature procedures [25–29] in our laboratory. The progress of the reaction was monitored by thin layer chromatography using aluminium sheets silica gel 60 GF-254, CH₂Cl₂–10% MeOH as eluent. H NMR analysis was made at 400 MHz and in D₂O. The maximum water contents of the liquids were determined using a 756 Karl Fisher coulometer.

2.1. Luminiscent bacteria acute toxicity test

The toxicity of the aqueous samples with the different ILs was determined by means of a bioassay following the standard Microtox test procedure (ISO 11348-3, 1998) based on the decrease of light emission by *P. phosphoreum* resulting from its exposure to a toxicant. The more toxic the sample is, the greater is the percent light loss from the test suspension of luminiscent bacteria. Bacterial bioluminescence has proved to be a convenient measure of cellular metabolism and consequently, a reliable sensor for measuring the presence of toxic chemicals in aquatic samples. Strain 11177 was originally chosen for the acute and chronic tests because it displays a high sensitivity to a broad range of chemicals.

A Microtox[®] M500 Analyzer (Azur Environmental) was used. The inhibition of the light emitted by the bacteria was measured after 15 min contact time. The IC₅₀ is defined as the ratio of the initial volume of sample (V_S) to the one yielding, after the required dilution, a 50% reduction of the light emitted by the microorganisms (V_F). Therefore the IC₅₀ is related to the dilution of the sample required to achieve a 50% of light emission reduction. From IC₅₀, the toxicity units of the wastewater are calculated as:

$$TU_{50} = \frac{100}{IC_{50}} \quad (1)$$

The EC₅₀ is defined as the effective nominal concentration of the toxic chemical (in mg/L) that reduces the intensity of light emission by 50%. Therefore, the IC₅₀ corresponding to a sample containing only one toxicant *i*, in concentration C_{*i*}, will correspond to:

$$IC_{50} = \frac{100C_i}{EC_{50}} \quad (2)$$

By using Eq. (2), the EC₅₀ of each ionic liquid is obtained from the IC₅₀ value of the aqueous sample containing the ionic liquid in a known amount C_{*i*}. A pH readjustment in order to prevent the pH effect before measuring the toxicity was not necessary because the pH of the samples ranged between 6.5 and 7. All the chemicals used in the toxicity test were purchased from

Sigma–Aldrich and the microorganisms were Microtox[®] Acute Reagent supplied by I.O. Analytical.

2.2. Biochemical oxygen demand

The biochemical oxygen demand of the aqueous sample containing the ionic liquid and/or other carbon source (glucose) was evaluated using a respirometric BOD measurement system (OxiDirect[®] by Lovibond). Biochemical oxygen demand is usually defined as the amount of oxygen (mg/L) required by bacteria while stabilizing decomposable organic matter under aerobic conditions. The term “decomposable” may be interpreted as meaning that the organic matter can serve as food for the bacteria, and energy is derived from its oxidation.

Solutions containing 100 mg/L of ionic liquid and/or 100 mg/L of glucose were prepared, in aerated media, pH of the liquid samples being in the range of 6.5–7.

A volume of 244 mL of each solution was then inoculated with 1 mL of an effluent collected from a biological reactor of a wastewater treatment plant and each well-mixed solution was dispensed into a series of BOD bottles (500 mL of volume) and incubated at 20 ± 1 °C in the dark for 5 days, obtaining the BOD₅ value.

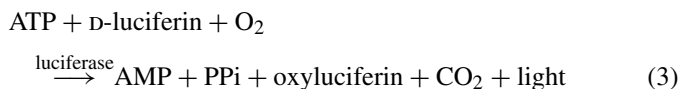
2.3. Glucose and ionic liquid measurement

D-Glucose was measured by UV-method using an enzymatic kit (Cat. No. 10 716 251 035 BOEHRINGER MANNHEIM / R-BIOPHARM). D-Glucose is phosphorylated to D-glucose-6-phosphate (G-6-P) in the presence of the enzyme hexokinase (HK) and adenosine-5'-triphosphate (ATP) with the simultaneous formation of adenosine-5'-diphosphate (ADP). In the presence of the enzyme glucose-6-phosphate dehydrogenase (G6P-DH), G-6-P is oxidized by nicotinamide adenine dinucleotide phosphate (NADP) to D-gluconate-6-phosphate with the formation of reduced nicotinamide adenine dinucleotide phosphate (NADPH). The amount of NADPH formed in this reaction is stoichiometric to the amount of D-glucose. The increase in NADPH is measured by means of its light absorbance at 334 nm. The test combination contains a bottle A, with approx. 7.2 g powder mixture, consisting of triethanolamine buffer, pH approx. 7.6; NADP, approx. 110 mg; ATP, approx. 260 mg; magnesium sulfate; a bottle B, with approx. 1.1 mL suspension, consisting of hexokinase, approx. 320 U; glucose-6-phosphate dehydrogenase, approx. 160 U. Bottle A content is dissolved with 45 mL redistilled water. Into the cuvettes, 1 mL of solution A with 2 mL of redistilled water and 0.020 mL of solution B is pipetted for the blank. For measurement of glucose in samples, 0.9 mL of solution A + 0.100 mL of sample is pipetted in the cuvette with 2 mL of redistilled water and 0.020 mL of solution B. Calibration of the absorbance measurements was made by using glucose standard solution from 10 to 100 mg/L.

Ionic liquid concentration in the aqueous solution was measured by UV at 210 nm. All the ILs tested showed a maximum at this wavelength. Calibration for each IL was made by preparing IL standard aqueous solution from 10 to 100 mg/L.

2.4. ATP determination

ATP in living cells was measured by a bioluminescence method using the ATP Biomass Kit (Biothema). All cells contain ATP, which plays the role of energy currency between different cellular processes. When cells die of natural causes, ATP is normally degraded. The intracellular concentration of ATP is carefully regulated to similar levels in all types of cells. ATP is therefore a good estimate of the total intracellular volume. Before the assay, ATP is released from the cell using the extractant B/S included in the commercial Kit cited. ATP is assayed using ATP reagent HS (highly sensitive) having a detection limit of 10^{-17} mol corresponding to five bacterial cells. The following reaction takes place:



The intensity of the light is proportional to the amount of ATP and it is measured in a luminometer (Optocomp I, MGM Instruments).

3. Results and discussion

3.1. Toxicity assessment

For all substances tested concentration–response curves were obtained and EC₅₀ values (μmol/L) were calculated. A typical plot of inhibition of the luminescence vs. the IL concentration (as dilution ratio $V_S/V_F \times 100$) is shown in Fig. 1.

The obtained EC₅₀ values at 15 min are given in Table 2, expressed as the corresponding logarithm, the confidence interval of the lognormal regression also being included. Literature EC₅₀ values for imidazolium ionic liquids are also given in Table 2 as well as the EC₅₀ values for typical volatile organic compounds. Notice that the EC₅₀ values obtained in this work are similar to the values found in literature for butyl, hexyl and octyl R₂ substituents. The methyl (C₁mimMSO₄) and ethyl (C₂mimESO₄) R₂ substituent were not previously analyzed in literature, the results obtained being consistent with the rest of the EC₅₀ values.

The data in Table 2 shows that the measured EC₅₀ values vary between 58,000 and 5 μM (corresponding logarithm 4.76 and 0.70, respectively) depending on their chemical structure. On the other hand, the influence of the anion on the EC₅₀ value is minimal taking into account the slight differences found for C_jmim (j = 6, 8) when counter ions Cl[−] or PF₆[−] are tested (log values of EC₅₀ being 2.18 and 2.11 for C₆mimCl and C₆mimPF₆, respectively; log values of EC₅₀ being 0.94 and 0.70 for C₈mimCl and C₈mimPF₆, respectively). The small effect of the anion on toxicity of imidazolium compounds has also been noticed by other authors [16–18] as can be deduced from the results in Table 2.

On the other hand, a longer chain length of the alkyl substituent R₂ results in a remarkable increase of the acute toxicity, obtaining a lower EC₅₀ value. The linear regression analysis of the log EC₅₀ vs. the alkyl chain length of R₂ has been carried

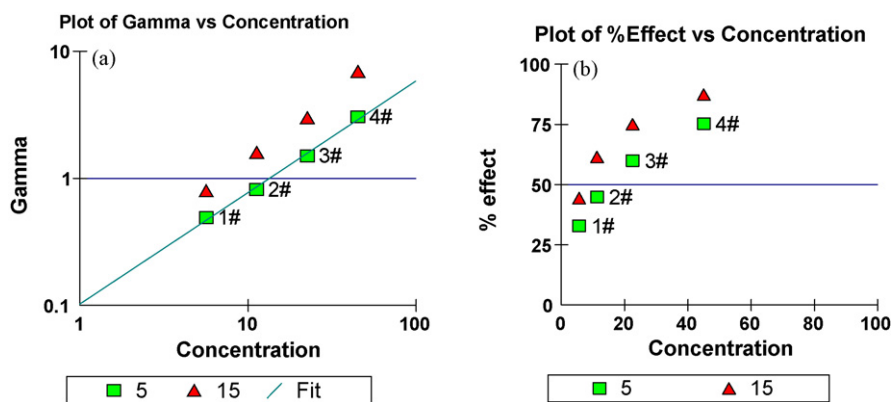


Fig. 1. Example of an inhibition plot of the luminescence vs. the concentration of the ionic liquid. The IL used was ClC_6mim at $C_i = 430 \text{ mg/L}$. (a) Gamma vs. concentration, as $(V_S/V_F) \times 100$ and (b) % effect vs. concentration.

Table 2
EC₅₀ values and confidence intervals for the tested ionic liquids

IL	log EC ₅₀ (μM), this work	log EC ₅₀ (μM) [18]	log EC ₅₀ (μM) [16]	log EC ₅₀ (μM) [17]	Common VOCs	log EC ₅₀ (μM) [14]
C ₁ mimMSO ₄	>4.76				Methanol	5.55–7.00
C ₂ mimESO ₄	4.02 ± 0.14				Ethanol	5.70–6.08
C ₄ mimCl	3.39 ± 0.15	3.71 ± 0.13	3.34 ± 0.13		2-Propanol	5.77–5.84
C ₄ mimPF ₆			3.07 ± 0.29		Acetonitrile	5.77
C ₄ mimBr		4.01 ± 0.05	3.27 ± 0.09	3.07 ± 0.03	Acetone	5.35–5.70
C ₄ mimBF ₄			3.10 ± 0.17	3.55 ± 0.04	Dicloromethane	4.07–4.53
C ₆ mimCl	2.18 ± 0.09		2.32 ± 0.16		Chloroform	3.55–4.32
C ₆ mimPF ₆	2.11 ± 0.11		2.17 ± 0.06		Benzene	1.41–3.12
C ₆ mimBr		1.42 ± 0.10			Phenol	2.35–
C ₆ mimBF ₄				3.18 ± 0.03	Toluene	2.29–2.64
C ₈ mimCl	0.94 ± 0.14		1.19 ± 0.11			
C ₈ mimPF ₆	0.70 ± 0.16		0.95 ± 0.12			
C ₈ mimBr		0.63 ± 0.06				
C ₈ mimBF ₄				1.41 ± 0.07		

out and the following equation has been obtained in this work:

$$\log \text{EC}_{50} = 5.33 - 0.549n_{\text{CR}_2} \quad (4)$$

This relationship has a similar slope with the number of carbons of the alkyl substituent in R₂ (n_{CR_2}) as the one obtained by Ranke et al. [17] analyzing $n_{\text{CR}_2} \geq 4$.

$$\log_{10} \text{EC}_{50} = +6.65 - 0.66n_{\text{CR}_1} - 0.57n_{\text{CR}_2} \quad (5)$$

n_{CR_1} being the number of carbons of the alkyl substituent in R₁.

In Fig. 2, the log EC₅₀ values are plotted vs. the alkyl chain length of R₂. The figure shows that a linear relationship is obtained for the interval $1 \leq n_{\text{CR}_2} \leq 8$. The validity of this relationship does not depend on the anion, and thus can be used to predict the toxicity of the R₂mim ILs.

3.2. Biodegradability

The 5 days BOD was measured for six samples containing 100 mg/L of D-glucose. An amount of 100 mg/L of IL was added in four of these D-glucose solutions. Tested ILs for this analysis have been C₂mimCH₃CH₂SO₄, C₄mimCl, C₆mimCl, C₈mimCl. They were selected in order to analyze the influence

of the alkyl chain length in the BOD₅. The experiments carried out for BOD₅ measurements as well as the obtained BOD₅ values are shown in Table 3. All the samples yield similar values of BOD₅, the average value being about 100 mg O₂/L. At the experimental conditions used, the expected BOD₅ of a sample

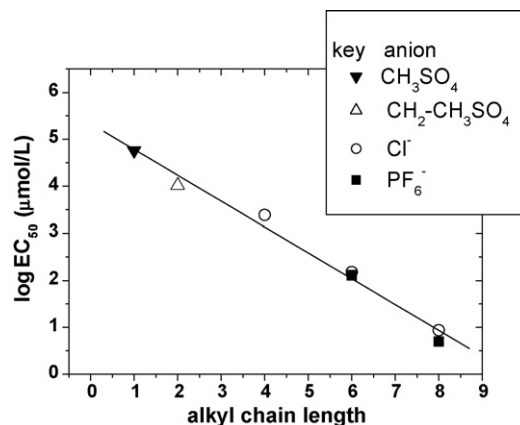


Fig. 2. Effect of the anion and alkyl chain length on the acute toxicity (Microtox[®]) for 1-alkyl-3-methylimidazolium ionic liquids.

Table 3
Biodegradability and microorganism inhibition for ionic liquid-D-glucose aqueous solutions (bioluminescence assay used for ATP measurements)

Sample	$t = 0$ days		$t = 5$ days				$t = 10$ days	
	D-Glucose (mg/L)	IL type	IL (mg/L)	BOD ₅ (mg O ₂ /L)	D-Glucose (mg/L)	IL (mg/L)	Light emission	IL (mg/L)
1	100	–	–	98	≈0	–	2.8×10^7	–
2	100	–	–	106	≈0	–	2.7×10^7	–
3	100	C ₂ mimC ₂ H ₅ SO ₄	100	93	≈0	89	2.1×10^7	90
4	100	C ₄ mimCl	100	101	≈0	98	2.2×10^7	97
5	100	C ₆ mimCl	100	100	≈0	99	1.3×10^7	98
6	100	C ₈ mimCl	100	102	≈0	95	1.2×10^7	96

containing 100 mg/L of D-glucose as carbon source was calculated as 106 mg O₂/L. Therefore, a higher value of BOD₅ than 100 mg O₂/L would be expected from samples containing IL as additional carbon source.

Final concentration of D-glucose and IL in the samples were analyzed, using an enzymatic kit or by UV-spectrometry, respectively. Values obtained are also given in Table 3. D-Glucose has been almost totally consumed in all the samples while the ionic liquid remains always in a concentration close to the initial one. Thus, all the ionic liquids tested were poorly biodegradable in presence of glucose. As the living organisms inoculated in this assay consumed only this monosaccharide and not the ionic liquid, the almost constant BOD₅ (100 mg/L) obtained is consistent.

Since this preference for the glucose was unsurprising, it was analyzed if the microorganisms would consume the IL in absence of another source of carbon. For this, the aqueous samples after the BOD₅ measurements, free of glucose, containing only the microorganism and the ionic liquids, were stirred in aerobic atmosphere for 5 days more at 20 °C in the dark. Remaining concentration of the ionic liquid at this final time of 10 days in contact with the bacteria was determined and values obtained are summarized in Table 3. The amounts of ionic liquid in the solutions after 10 days were almost the same as the initial ones and those obtained after 5 days. In conclusion, the poor biodegradability of ILs in presence and absence of other carbon source is confirmed.

3.3. Influence on the biological activity

To examine if the presence of an ionic liquid produce cellular stress on the microorganism in the aqueous phase, a bioluminescence assay have been used to measure the ATP in the samples after BOD₅ experience. The light emitted (proportional to the ATP in the samples) is given in Table 3. Control samples are set like those containing only glucose in the BOD₅ measurements. In Table 3, the light emission values obtained in the samples that contained ILs were lower than those obtained in the control samples. Therefore, the aqueous samples with ionic liquid added yield a lower living cells concentration than the control samples. Because all the samples were inoculated with the same type and amount of microorganisms, it can be deduced that a lower microorganism growth takes place if the microorganism is inoculated in a medium containing an ionic liquid. Again, the shorter the chain length of side R₂, the lower the toxic effect.

The main limitation of the Microtox test is that the toxic effects noticed against the microorganism *V. fischeri* cannot be directly extrapolated to predict the toxicity effect against other organisms, as those living in the biological reactor of a wastewater treatment plant.

Nevertheless, it has been found that the IL compounds are not biodegradable and cause cellular stress, these results being in agreement with the high ecotoxicity values obtained for some ILs by using the Microtox test. We consider that the non-biodegradability related to the cellular stress of the microorganisms in the activated sludge is an important finding not previously described in literature.

Moreover, the same trend is noticed for the effect of the chain length of the alkyl substituent R₂ on both cellular stress and ecotoxicity, thus the utility of the Microtox test to elucidate the potential toxicity of a compound can be validated.

4. Conclusions

It has been shown that imidazolium based ionic liquids have a wide range of toxicities in short bioassay used (Microtox®). In general, their toxicity (EC₅₀ value) correlates directly with the length of the *n*-alkyl substituent in the methyl imidazolium cation while the anion has a low effect on this EC₅₀ value. These new solvents can be more toxic towards cells than conventional solvents and this must be taken into consideration with regard to their fate and persistence in the environment. In fact, the ILs tested were poorly biodegradable and the microorganisms do not consume them as carbon source. If the aqueous sample contains an ionic liquid it is certain that the microorganisms used in the BOD₅ determination suffer a cellular stress: a lower growth of the microorganism inoculated at the same operational conditions occurs when there is an ionic liquid in the media.

The relatively high solubilities of the ILs in the water phase, the low EC₅₀ values obtained for some of them and their poor biodegradability (that makes them persistent pollutants) have important environmental consequences and should be taken into account for the design of processes that use ILs. To avoid the potential contamination of the aqueous phase with ILs, several strategies should be planned. Firstly, it is important to improve the processes, minimizing the IL leaches to the aquatic media. Furthermore, downstream separation step must be required at the end of these processes to remove the ILs from wastewater streams.

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